

New policy directions for global pond conservation

Short running title: Global pond conservation

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41

42 **Abstract**

43 Despite the existence of well-established international environmental and nature conservation policies
44 (e.g., the Ramsar Convention and Convention on Biological Diversity) ponds are largely missing from
45 national and international legislation and policy frameworks. Ponds are among the most biodiverse
46 and ecologically important freshwater habitats, and their value lies not only in individual ponds, but
47 more importantly, in networks of ponds (pondscapes). Ponds make an important contribution to
48 society through the ecosystem services they provide, with effective conservation of pondscapes
49 essential to ensuring that these services are maintained. Implementation of current pond conservation
50 through individual site designations does not function at the landscape scale, where ponds contribute
51 most to biodiversity. Conservation and management of pondscapes should complement current
52 national and international nature conservation and water policy/legislation, as pondscapes can provide
53 species protection in landscapes where large-scale traditional conservation areas cannot be established
54 (e.g., urban or agricultural landscapes). We propose practical steps for the effective incorporation or
55 enhancement of ponds within five policy areas: through open water sustainable urban drainage
56 systems in urban planning, increased incentives in agri-environment schemes, curriculum inclusion in
57 education, emphasis on ecological scale in mitigation measures following anthropogenic
58 developments, and the inclusion of pondscapes in conservation policy.

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60 **Key words:** biodiversity; ecosystem services; freshwater policy; international nature conservation;
61 landscape-scale; pond networks; small waterbodies.

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67 **Introduction**

68 Longstanding international environmental and nature conservation policies (such as the Ramsar
69 Convention, the Convention on Biological Diversity, and the European Water Framework Directive)
70 are important for protecting species and habitats, in the face of growing anthropogenic pressures
71 (Dudgeon *et al.* 2006). Despite this, the number of threatened species listed on the IUCN Red List
72 continues to increase, human-dominated lands (urban, agricultural) continue to replace natural lands
73 (Decker *et al.* 2016), and a number of key terrestrial and freshwater habitats continue to be
74 overlooked by policy makers. Ponds, defined in the UK and most of Europe as lentic waterbodies <2
75 ha in area (Williams *et al.* 2010), and pondscapes, defined as a network of ponds and their
76 surrounding terrestrial matrix (Fig. 1), are one such historically neglected habitat. Recently, there has
77 been a significant increase in recognition of the importance of ponds and pondscapes to biodiversity
78 and ecosystem services by scientific and non-scientific communities. Yet these small waterbodies
79 remain largely outside the remit of international, and in many cases national, conservation and
80 environment legislation.

81

82 **Current conservation status of ponds**

83 Ponds are abundant across the globe (c. 500 million ponds and lakes are estimated to exist worldwide;
84 Holgerson & Raymond 2016) and are critically important for ecology and society. Recent evidence
85 indicates that pondscapes support high biodiversity (see The Pond Manifesto: EPCN 2008), and
86 contribute disproportionately more to catchment aquatic biodiversity than larger and more widely
87 studied freshwater bodies such as lakes and rivers (Davies *et al.* 2008b). Furthermore, ponds provide
88 essential habitat for many nationally and internationally rare and threatened species and are important
89 refuges in urban and agricultural landscapes (Davies *et al.* 2008b; Chester & Robson 2013). The
90 significant contribution of pondscapes to local and regional aquatic biodiversity can be attributed to
91 (i) the small catchments of individual ponds, resulting in idiosyncratic environmental conditions and
92 habitat complexity, leading to landscape scale habitat heterogeneity (Davies *et al.* 2008b), (ii) the
93 value of anthropogenic ponds (e.g. farm ponds) for increasing the area of freshwater habitat available

94 for wildlife, and (iii) the provision of refuge habitats for aquatic communities, especially where
95 natural wetlands have been largely converted into farm ponds or paddy fields (Takamura 2012;
96 Chester & Robson 2013). Ponds also play an important role in supporting semi-aquatic and terrestrial
97 flora and fauna, for example, agricultural areas that contain ponds support higher richness and
98 abundance of terrestrial species than agricultural areas without ponds (syrphids and bees: Stewart *et*
99 *al.* 2017; birds: Davies *et al.* 2016).

100

101 Ponds are increasingly recognised for the important ecosystem services they provide to society
102 including flood alleviation, storage of urban storm water, the supply of irrigation water (Takamura
103 2012), and nutrient or pesticides removal from water. Ponds have been estimated to sequester a
104 similar amount of carbon to the world's oceans (Downing *et al.* 2008), and may help mitigate the
105 effects of urban heat islands (Coutts *et al.* 2012). These small waterbodies also provide considerable
106 amenity, providing social and cultural benefits including improved physical and mental wellbeing and
107 increased awareness of biodiversity and nature conservation (Lundy & Wade 2011).

108

109 Despite these benefits, current national and international environmental legislation and management
110 strategies are almost exclusively focussed on large waterbodies, with ponds being harder to
111 characterise, evaluate, monitor, and protect (see Table S1). In Europe, the EU Water Framework
112 Directive (WFD) was implemented to protect and improve the water quality of all freshwaters and
113 shallow coastal water. Yet in practice the WFD only covers rivers and standing waterbodies ≥ 50 ha,
114 while excluding the vast majority of small wetlands and lentic waterbodies (Hassall *et al.* 2016). It has
115 been suggested that the broad catchment-scale measures implemented under the WFD for larger
116 waterbodies should also protect smaller waterbodies, but there is little evidence for this (Biggs *et al.*
117 2016). Nature conservation organisations have been quicker to identify the value of ponds for
118 biodiversity, and as a result, nature conservation legislation at a European scale is currently the most
119 important tool for protecting pond habitats and their biota (Hassall *et al.* 2016). The Habitats Directive
120 provides protection as priority habitat to only a few pond types (e.g., Mediterranean temporary ponds,

121 natural dystrophic lakes and ponds) and some pond associated species (e.g. Great Crested Newt,
122 *Triturus cristatus*) (EC 1992). However, the scale at which pond conservation is applied currently (i.e.
123 individual site designation) is not the scale at which ponds contribute most to biodiversity (pondscape
124 scale; Hill *et al.* 2016). Furthermore, as significant advances in knowledge of pondscales occurred
125 after the implementation of the Habitats Directive and WFD, there remain significant gaps in the
126 protection that these directives provide to pondscales and their biota (Biggs *et al.* 2016).

127

128 Similarly, in North America and Australia, pond habitats in general do not receive direct legislative
129 protection despite the Clean Water Rule 2015 of the Clean Water Act in the USA (Department of
130 Army, Corps of Engineers and US Environmental Protection Agency 2015) and the national-scale
131 Environment Protection and Biodiversity Conservation Act 1999 in Australia (Act 1999). The latter
132 includes an inventory of > 900 ‘nationally-important’ wetlands that are protected based on meeting \geq
133 1 of 6 criteria. However, these criteria focus on attributes of single wetlands, ignoring their landscape
134 contexts, and few small waterbodies are specifically designated.

135

136 In Asia, international legislation specifically targeted for conservation of pond habitats is largely
137 lacking. A possible reason for this is a lack of holistic biodiversity or water quality surveys of ponds
138 in international or national monitoring programmes. In Japan, the Ministry of Land, Infrastructure,
139 Transport and Tourism (1990-) conducts nation-wide and long-term censuses of water quality and
140 biodiversity in rivers and impounded reservoirs, local governments (1970-) monitor water quality of
141 lakes, and the Ministry of the Environment (monitoring site 1000 programme; 2009-) evaluates
142 biodiversity of some 20 wetlands or lakes. Although action plans for wetland/pond conservation exist
143 in conjunction with The National Biodiversity Strategy of Japan 2012-2020, comprehensive or long-
144 term water quality and biodiversity monitoring data are lacking. In India, almost all ponds and
145 pondscales are excluded from environmental and nature conservation legislation. Some large
146 wetlands in India receive legislative consideration from the Wetlands (Conservation and
147 Management) Rules under the Environment Protection Act 1986 and the National Environment Policy

148 2006, which regulates the activities that can be undertaken within these wetlands (Sundar & Kittur
149 2013). A few ponds within protected areas also receive some protection under the Wildlife Protection
150 Act 1972. However, there have been heated debates between civil society and central governments
151 because of restrictions on activities in these wetlands (although in Sept 2017, wetland conservation
152 and management legislation in India was amended, potentially excluding certain categories of human-
153 made water bodies). Legislation for the sustainable management of ponds in India is difficult as
154 irrigation departments, fisheries departments and district councils often manage them jointly.

155

156 Ponds may receive indirect protection through other legislation. For example, South Africa's National
157 Biodiversity Strategy and Action Plan (2005; NBSAP) and Namibia's National Biodiversity Strategy
158 and Action Plan (2013; NNBSAP) aim to integrate terrestrial and aquatic management to minimise
159 the impacts of processes that threaten biodiversity, to enhance ecosystem services and, to improve
160 social and economic security. In such water scarce countries, pond security is thereby embedded in
161 policies aimed at water security, biodiversity conservation and resilience without specifically referring
162 to the thousands of water retention ponds throughout these nations. Importantly, both these southern
163 African NBSAPs emphasise not only hydrology, resilience and sustainability, but also the importance
164 of conserving the rich heritage of endemic species.

165

166 At the international level, the Ramsar Convention, signed by 169 countries, ensures that key wetland
167 (Ramsar) sites of international importance are protected (Ramsar 2016), with many encompassing
168 large numbers of ponds. Other international initiatives, particularly the Convention on Biological
169 Diversity, have stimulated development of the international partnership for 'Satoyama Initiatives' to
170 promote the sustainable management and use of natural resources that benefit society and
171 biodiversity, partially incorporating pondscapes (Bélair *et al.* 2010). In addition, there are a few
172 national-scale policies that provide protection for ponds, such as the UK's recognition of 'Priority
173 Habitats' and 'Priority Species' for site-specific and species-specific conservation and management
174 (JNCC & Defra 2012). However, most ponds and pondscapes fall outside of contemporary nature

175 conservation policy. This has arisen largely from a lack of recognition and poor understanding of the
176 importance of pondscapes for sustaining local and regional biodiversity at a policy/management level.
177 In arid and semi-arid countries like Namibia and South Africa, where ponds are included in regional
178 plans, the focus is to protect water resources and promote hydrological cycles rather than pond
179 biodiversity *per se*. The reality of global freshwater conservation is nuanced and complex,
180 incorporating a range of political issues (e.g., definitions of different freshwater habitats, top-down vs.
181 bottom-up management), social issues (property and societal rights), and economic issues (economic
182 development vs environmental conservation, and cost effectiveness of management) (Calhoun *et al.*
183 2014). However, there remain significant opportunities for the inclusion of ponds and pondscapes in
184 international and national conservation and policy frameworks.

185

186 **Opportunities for pond conservation, supported by science**

187 *Patch-network conservation*

188 Knowledge regarding the value of pondscapes, even in human-dominated environments (Hill *et al.*
189 2016), now provides clear empirical evidence and support for their inclusion in environment and
190 nature policy frameworks. Conceptual advances in ecological research from ‘corridors’ to
191 ‘connectivity’ has provided critical scientific evidence to underpin the development of practical
192 conservation strategies across landscapes. Groups of small habitats generally provide as high (or
193 higher) conservation value than a single large habitat of equal area (Fahrig 2017). Several studies
194 have shown that networks of smaller ponds support higher taxonomic richness and conservation value
195 than one large pond (Martinez-Sanz *et al.* 2012; Oertli *et al.* 2002). In addition, higher pond density is
196 associated with greater species richness in UK urban ponds (Gledhill *et al.* 2008). This suggests that
197 the current legislative focus on large, contiguous habitats at the exclusion of small habitats is
198 potentially misguided (e.g. EU Natura 2000 network: European Commission 2008). Further, applying
199 patch-network conservation exclusively at large spatial scales can be ecologically ineffective, missing
200 local scale biodiversity hotspots, particularly in human-dominated landscapes. One way to improve
201 the effectiveness of landscape conservation is to incorporate networks of smaller freshwater habitats

202 (\approx ponds), alongside large-scale habitat networks. For example, in the context of widespread
203 agro-forestry in South Africa, large-scale ecological networks (ENs) of remnant land within
204 agroforestry landscapes have been set aside to mitigate the effects of agro-forestry. These ENs are
205 rich in natural and artificial ponds and are similar in biodiversity value to those in neighbouring
206 protected areas (Pryke *et al.* 2015). However, urban and agricultural landscapes often represent
207 barriers (e.g., roads) for the dispersal and colonisation of pond biota. Managing ponds would
208 increase focus on management actions that increase connectivity between ponds, especially for native
209 species migration between ponds (e.g. culverts beneath roads, restoring drains as streamlines with
210 fringing vegetation). Consideration of ponds favours landscape complementation because they
211 encompass a variety of habitat types (proximal terrestrial and aquatic habitat) for many species to
212 complete their life histories (Pope *et al.* 2000). Furthermore, conservation of ponds facilitates
213 connectivity and dispersal, particularly in agricultural landscapes, acting as stepping stones between
214 larger protected freshwater habitats, thereby increasing the effectiveness of conservation measures at
215 larger spatial scales (Kukkala *et al.* 2016; Pryke *et al.* 2015).

216 In human-dominated landscapes, many ponds and ponds are located on private land and if faced
217 with the prospect of mandatory conservation initiatives, it may be financially and logistically easier
218 for landowners to remove (i.e. drain or infill and build over) ponds given their small size (Calhoun *et al.*
219 *et al.* 2014). In some agricultural and urban landscapes where private ownership of ponds is high,
220 environment and conservation legislation may need to be flexible and designed to allow
221 environmentally friendly farming, forestry, fisheries, ecotourism and/or urban development to ensure
222 the persistence and protection of ponds, while not overly restricting local economic activities (Usio &
223 Miyashita 2014). For example, most ponds in Japan are used for irrigation for rice farming and
224 form a part of Satoyama, a landscape mosaic of paddy fields, dry cropland, farm ponds, grassland,
225 secondary forests, streams and villages. Given that the biodiversity of Satoyama is maintained through
226 traditional farming, forestry and fishing activities, moderate levels of human activities are encouraged
227 to maintain indigenous biodiversity as well as to sustain the local economy (Takeuchi 2010).
228 Furthermore, to raise public awareness of the value of multifunctionality in agricultural areas, the

229 FAO (Food and Agriculture Organization of the United Nations) designates regions with traditional
230 agriculture, indigenous culture, scenic landscape and sustainable use of natural resources as Globally-
231 Important Agricultural Heritage Systems (GIAHS 2017). Biodiversity conservation in urban areas
232 presents a number of challenges associated with development. However, ponds are increasingly
233 recognised for the ecosystem services they provide in cities. In some new urban developments,
234 stormwater / groundwater recharge ponds have been created. These provide some 'natural' habitat,
235 offsetting pond loss, and maintaining biodiversity in new developments (Hassall & Anderson 2015).

236

237 *Monitoring ecological condition*

238 Ponds and pondscapes are rarely monitored in a systematic manner because of the resource and
239 logistical implications for protecting these abundant waterbodies. Other monitoring options are
240 possible, such as the use of sentinel sites that can be monitored over long time-periods, citizen-science
241 based monitoring projects, or environmental DNA techniques that may facilitate rapid and effective
242 assessment of pond biodiversity and presence of protected species (Biggs *et al.* 2016). Monitoring
243 approaches need to be further refined to provide rapid, low cost assessments of the environmental and
244 biological quality of ponds to guide conservation management (Rosset *et al.* 2013). This is currently
245 being implemented in South Africa using a Dragonfly Biotic Index, which can be applied to small
246 pond environments as well as other freshwater systems (Samways & Simaika 2016). Monitoring a
247 charismatic taxon like dragonflies, which may also act as an umbrella for many other taxa, makes data
248 collection more feasible, especially as citizen scientists can be readily engaged. In India, the
249 identification of ponds and wetlands through the development of the Wetland Atlas (Bassi *et al.* 2014)
250 provides significant opportunities for the periodic monitoring of pondscapes using remotely sensed
251 data and citizen scientists.

252

253 Ponds provide frequent opportunities for citizens to engage in conservation and habitat management
254 activities, especially when linked to education or enjoyment of wildlife through dedicated trails

255 (Willis & Samways 2013). Given the inadequate funding levels for global biodiversity conservation
256 (Waldron *et al.* 2013), there is increasing reliance on agencies such as environmental charities to act
257 as intermediaries among government policy makers, stakeholders and the public to realise the
258 aspirations of conservation initiatives. The development of a forum which connects stakeholders such
259 as scientists, landholders, citizens, environmental groups/agencies and policy makers may facilitate
260 pond conservation. Such a forum should provide digital and/or physical space for dialogue among
261 groups, make scientific findings accessible to resource managers, stakeholders and citizens, provide
262 training in pond monitoring, and facilitate the development of conservation initiatives that are robust,
263 innovative and accessible for all groups (Calhoun *et al.* 2014).

264

265 **Policy-based recommendations**

266 Sufficient research now exists to underpin policy recommendations for ponds. There is an ecological
267 need for the Ramsar Convention, the Convention on Biological Diversity and other international
268 environmental legislation (e.g., the WFD in Europe) to now explicitly recognise ponds. Below,
269 we provide recommendations on how ponds, ponds, and their ecosystem services should be
270 incorporated into policy:

271 (1) Environmental context – Given that ponds often occur in networks linked by important terrestrial
272 habitats, identifying groups of important sites as management units (recommended by the WFD; EC
273 2003) will be logistically easier and more cost effective than monitoring/protecting individual ponds.
274 Defining ponds as management units increases opportunities to monitor ponds over wider areas
275 and to identify objectives for each pond (Biggs *et al.* 2016). In addition, requiring permits for
276 modifications (positive or negative) of ponds provides a policy tool that can consider the role of each
277 pond within the pond, and would require applicants to maintain/enhance a pond's capacity to
278 sustain native biodiversity within the pond. Local government or non-government environmental
279 organisations would be well-placed to implement pond management units and permits.

280 (2) Urban planning – Planning regulations can be adapted to prioritise open water sustainable urban
281 drainage systems alongside other nature-based solutions (Dadson *et al.* 2017). Mitigation for pond
282 loss during development should be based on pondscape-scale considerations rather than individual
283 habitat creation. Also, during urban development, there should be a focus on zero ecological loss, as
284 opposed to zero habitat loss, and ponds could form a key part of this strategy. Under some conditions,
285 stormwater ponds can support significant biodiversity (Hassall & Anderson, 2015), especially where a
286 treatment train of clean water ponds (e.g. receiving roof water) is initially separated from ponds
287 receiving contaminated water (e.g. from roads or vehicle parks). Diverting runoff water that would
288 otherwise flow directly to lakes or rivers, into such ponds, could increase pond density and
289 biodiversity in urban areas as well as help mitigate flooding and retain pollutants.

290 (3) Flood management – The current trend towards natural flood management provides an opportunity
291 for policies to incorporate pondscales. Ponds can be easily integrated into open water flood storage
292 strategies because small waterbodies may pose fewer logistical issues than larger ones, yet hold an
293 equivalent volume of water. It may be also relatively easy to integrate numerous small ponds into
294 urban or rural land management schemes, such as the “sponge city” concept currently being adopted
295 in China (Liu *et al.* 2017).

296 (4) Agriculture – Financial incentives are sometimes provided (e.g., under the EU agri-environment
297 schemes) for the maintenance of individual farmland ponds of significant biodiversity value (Attwood
298 *et al.* 2009; Davies *et al.* 2008a). These incentives could be modified to ensure that the protection and
299 creation/restoration of pond networks is rewarded at a rate greater than the sum of the individual
300 ponds, provided collaborative agreements could be made between multiple landowners.

301 (5) Education - Opportunities may exist for “pond schools” which parallel “forest schools” in their
302 focus on nature as a core of education (Austin *et al.* 2016). Many schools in urban or rural landscapes
303 could make greater use of nearby ponds to provide enhanced pedagogical and health benefits. In
304 addition, as part of the increased focus on nature play and kitchen gardens in schools, ‘frog ponds’
305 could be constructed to provide these benefits to students and their communities. In human-dominated

306 landscapes, public awareness of pondsapces can be increased by designating globally or nationally
307 important pondsapces (through frameworks such as GIAHS 2017).

308

309 **Conclusion**

310 Current conservation policy is failing to preserve much of the aquatic biodiversity and ecosystem
311 services supported by pondsapces. For policy to be consistent with current scientific understanding,
312 pondsapces should be better integrated into national and international policy frameworks to maximise
313 opportunities for conserving and protecting biodiversity and ecosystem services. Although the
314 economic implications of new environmental policies will be contested in certain quarters, because of
315 their small size pondsapces may be easier to conserve and maintain than larger waterbodies. Moving
316 away from site-specific conservation to a strategy that conserves resilient landscapes, puts people at
317 the heart of the environment, and grows natural capital will promote biodiversity conservation
318 (Natural England 2016). An evidence-based conservation strategy that incorporates pondsapces into
319 policy frameworks will significantly improve existing legislation by protecting a valuable,
320 multifunctional habitat type that provides a solution to multiple complex societal challenges while
321 supporting and enhancing biological diversity.

322

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454 **Figure captions**

455 Fig. 1 - Groups of small ponds and surrounding habitats ('pondscapes') provide important ecosystem
456 services in human-dominated landscapes. A pondscape in (a) an agricultural landscape in the UK
457 (Leicestershire) and (b) an urban setting in Australia (Perth; providing important habitat for
458 amphibian metacommunities). Map data credit: Google Earth 2016.

459

(a)



(b)

